Bird Conservation International

www.cambridge.org/bci

Research Article

Cite this article: Solovyeva D, Lei J, Tian H, Fan R, Vartanyan S, Danilov G, Barykina D, Meng F, Lu C, Zeng Q, Lei G (2025). New information on the breeding and moulting ecology of the Eastern population of Lesser White-fronted Goose *Anser erythropus* from GPS/GSM tracking data. *Bird Conservation International*, **35**, e1, 1–12 https://doi.org/10.1017/S0959270924000285

Received: 21 January 2024 Revised: 22 August 2024 Accepted: 09 September 2024

Keywords:

Key moulting site; breeding propensity; site fidelity; nest success; timing of nesting

Corresponding author: Qing Zeng; Email: zengqing@bjfu.edu.cn

*These authors contributed equally to this work.

© The Author(s), 2025. Published by Cambridge University Press on behalf of BirdLife International.



New information on the breeding and moulting ecology of the Eastern population of Lesser White-fronted Goose *Anser erythropus* from GPS/GSM tracking data

Diana Solovyeva¹*, Jialin Lei²*, Haitao Tian³, Rong Fan², Sergey Vartanyan⁴, Gleb Danilov⁵, Daria Barykina¹, Fanjuan Meng⁶, Cai Lu², Qing Zeng² and Guangchun Lei²

¹Institute of Biological Problems of the North, Far East Branch, Russian Academy of Sciences, Magadan, Russia; ²Center for East Asian-Australasian Flyway Studies, Beijing Forestry University, Beijing, China; ³Jingzhou Municipal Bureau of Natural Resources and Urban-Rural Development, Hubei, China; ⁴North-East Interdisciplinary Scientific Research Institute n. a. N. A. Shilo, Far East Branch, Russian Academy of Sciences, Magadan, Russia; ⁵Peter the Great Museum of Anthropology and Ethnography, Russian Academy of Sciences, St. Petersburg, Russia and ⁶State Key Laboratory of Urban and Regional Ecology, Research Center for Eco-Environmental Sciences, Chinese Academy of Sciences, Beijing, China

Summary

The Eastern population of the Lesser White-fronted Goose (EPLWFG) Anser erythropus is shared between Russia and China. The summer range of the EPLWFG has been recognised as a continuous area extending from the Olenyok River in the west to the Anadyr River in the east and northwards from 64°N. The aim of this study was to provide information on breeding behaviour; nest-sites, nesting habitats, and time of nesting; nesting success; timing of summer movements including moult migration; moult timing, duration, and moulting habitats; site fidelity; and the effect of human presence. To accomplish this, we combined the results from field surveys with GPS/GSM tracking. A total of 30 summer tracks from 19 individual EPLWFG were analysed. We estimated breeding propensity in 93.8% of adult LWFG, and this factor did not seem to depend on breeding success in the previous season. Reproductive success was 13.3% in all nesting attempts. Non-breeders arrived three-week later and departed a week earlier. The EPLWFG are highly mobile during the summer. The core moulting site for the entire EPLWFG was discovered by this study and is located along the lower reaches of the San-Yuryakh and Kyuanekhtyakh rivers flowing towards the Omulyakhskaya Bay of the East Siberian Sea. The EPLWFG flightless period was 24.8 ± 2.8 days. A part of failured EPLWFG (43.7 %) migrated back to its early summer breeding/staging site after having completed moult. The strong site fidelity (100%) of adult birds to both nesting and moulting sites promotes the formation of local breeding populations, which could be considered conservation units if genetic studies support this differentiation. The EPLWFG selects the remotest and least human-accessible area for their remigial moult, and the main site was discovered with the help of tracking.

Introduction

The Eastern population of the Lesser White-fronted Goose Anser erythropus (hereafter EPLWFG), is shared between Russia, where the birds reproduce and moult, and China, where the majority winters (BirdLife International 2018; Cao et al. 2018). Far fewer birds winter in Japan, but the numbers have recently increased (Ao et al. 2020; Ikawa and Ikawa 2009). Recent studies using satellite tracking have revealed the wintering range, wintering habitats, and seasonal migrations of the EPLWFG (Ao et al. 2020; Lei et al. 2019a, b; Wang et al. 2012). The summer range has been recognised as a continuous area extending from the Olenyok River (119.2°E) in the west to the Anadyr River (174.8°E) in the east and northwards from 64°N, excluding the Arctic archipelagos (Tian et al. 2021). This finding contrasts with previous assumptions of a fragmented summer range (Cao et al 2018; Morozov 1995; Morozov and Syroechkovski 2002). The exclusion of Arctic archipelagos is based on a literature inventory with no LWFG ever reported outside the mainland of Eastern Asia. A study by Tian et al. (2021) does not support the overlap between Western Main and Eastern populations of the LWFG on their breeding grounds and suggests a gap between 103° and 119°E. The Western Main population winters around the Caspian Sea, in the Mesopotamian Lowland, and in south-eastern Europe, and the easternmost edge of the wintering range is on the coast of the Aral Sea (Aarvak and Øien 2018; Romanov and Pospelov 2010). The EPLWFG mainly winters in the middle and lower reaches of the Yangtze River in China (Wang et al. 2012). Thus, the two subpopulations are well separated in winter; their winter ranges lie over 5,000 km apart.

Being classified as "Vulnerable" by the International Union for Conservation of Nature (IUCN) (Birdlife International 2018), the LWFG is listed in the Red Data Book of Russian Federation (category "Endangered"; Morozov 2021) and in China, thus receiving a fully legal protection in both key states of the EPLWFG.

Ornithologists have divided the LWFG into three populations worldwide, i.e. the Fennoscandian (now only remains in Norway), Western Main, and Eastern Palearctic populations (Fox and Leafloor 2018). Between 1981 and 1999, restocking in Finland and Sweden involved captive breeding birds; however, the programme was interrupted because of a proposal to mix with Greater White-fronted Goose *Anser albifrons* (Andersson and Holmqvist 2010; Ruokonen et al. 2000). Recently, Swedish LWFG were found to be genetically distinct from the Fennoscandian and Western Main populations, with no evidence of interspecific introgression into the Swedish LWFG population (Molino et al. 2020).

Among the LWFG populations, the critically endangered Fennoscandian population (numbering a few dozen pairs; BirdLife International 2021) has been studied the most for breeding habitats, nesting ecology, reproductive success, moult migration and moult, and juvenile and adult survival rates (Aarvak and Øien 2003, 2008; Markkola and Karvonen 2020; Marolla et al. 2023; Øien and Aaevak 2008). This is followed by the Swedish population (fewer than 200 birds), in that nest-site fidelity, predation pressure, productivity, and moulting habitats have been studied (Andersson 2004). We typically lack detailed data on the breeding biology of both Western Main and Eastern populations of the LWFG. For the Western Main population, nesting habitats, and some aspects of nesting and postbreeding ecology have been described (Morozov 2021; Morozov and Kalyakin 1997; Rozenfeld et al. 2019) and the nesting association with Peregrine Falcon Falco peregrinus was proposed from observations of nests and newly hatched broods in the vicinities of peregrine nests on Yamal and Gydan Peninsula (Korobitsyn and Tyutenkov 2023; Pokrovskaya et al. 2023). Brood sizes and proportions of juveniles in brood-rearing flocks have been investigated on the Yamal Peninsula and the Polar Ural Mountains (Morozov and Syroechkovski 2002). A series of papers by Romanov (2005, 2008, 2015) described brood-rearing habitats, diet, and brood sizes on the lakes of Putorana Plateau. The feeding ecology of western LWFG was investigated in detail (Rozenfeld 2009).

In the EPLWFG breeding and moulting ecology is poorly documented relative to other populations. Few nests have ever been described (Andreev 2001; Degtyarev and Perfilyev 1996). Brood sizes and proportions of juveniles in local populations have been followed in Yakutia (Artiukhov and Syroechkovski 1999; Degtyarev and Perfilyev 1996; Egorov and Okhlopkov 2007) and Chukotka (Solovieva and Vartanyan 2011). LWFG summer records during the last decade are available from the Database of Anseriformes Air Surveys and Remote Tracking in Russia (http://rggsurveys.ru/).

Selection of remote and unpopulated rivers for brood rearing (and presumably nesting) was proposed for the EPLWFG (Egorov and Okhlopkov 2007; Solovieva and Vartanyan 2011). Modelling showed that human disturbance affects summer site suitability, with a decrease in species presence starting around 160 km from human settlements (Tian et al. 2021). However, the actual remoteness of breeding, moulting, and summering sites has never been measured. Standard numeric monitoring by ground counts is almost impossible for the EPLWFG (compared with other populations of the species), because birds of this population are able to avoid human presence by summering in the huge and remote areas of the forest, forest–tundra, and tundra of Eastern Siberia and the Far East of Russia. A vast majority of their summer sites are inaccessible for humans. Thus, it is not surprising

that knowledge of nesting habitats, breeding phenology, summer movements (including moult migration), moult duration, moulting habitats, nest success, site fidelity, and breeding propensity is still lacking for the EPLWFG. The best way to obtain this information is by the use of tracking techniques. The aim of this study was to provide data on breeding propensity; nest-sites, nesting habitats, and time of nesting; nest success; summer timing, summer movements including moult migration; moult timing, duration and moulting habitats; site fidelity; productivity; the effect of human presence on the selection of nesting and moulting sites. To accomplish this, we combined the results from field surveys with GPS/GSM tracking to generate information on the breeding and moulting ecology of the EPLWFG. Key moulting sites of the EPLWFG were discovered during this study.

Methods

Capturing and tracking methods

We received data for 11 birds of the EPLWFG that were caught during the winter of 2016/7 at East Dongting Lake, China (designated with the prefix BFUL, Supplementary material Table S1), and tagged with transmitters developed by the Hunan Global Messenger Technology Company (Chanhgsha City, China). These transmitters were programmed to record GPS position and speed. The backpack design transmitters measured 55 × 36 × 26 mm and weighed 22 g, approximately 1.6% of the bird's body mass, with a coordinate recording interval of two hours. All field methods used in this study were consistent with and approved by the Forestry Department of Hunan Province under a scientific research licence (No.11 Xiang Forest Protection 2014) (Lei et al. 2019b). The captured birds were not aged or sexed due to the field crew's lack of training. The authors have full access to the data set.

During the summering period, 10 adult LWFG of unknown sex were captured from brood-rearing flocks (nine birds) and nonbreeding flocks (one bird) along the Rauchua River (68.8°N 167.7° E) in July–August 2018 (designated with the prefix Lwfg; Table S1). Flightless birds were captured by driving them by motorboat from the water on to the riverbanks, where they were captured by hand. All birds were fitted with backpack solar-powered GPS/GSM transmitters developed by the Ornitela Company (Vilnius, Lithuania). These transmitters weighed 25 g, approximately 1.4% of the bird's body mass, and had a coordinate recording interval of 10 minutes (Ao et al. 2020). The backpacks in Teflon harnesses were placed over the wings and stitched with floss *in situ* on the bird, adjusted to breast size, with 1 cm freedom. The authors were granted access to the data set between 25 May and 10 October.

Track analyses

Each EPLWFG track was displayed as a separate layer in the QuantumGIS software, and each track was divided into migration (northbound and southbound), breeding, and moulting. Migration tracks were not utilised for this paper, except for the dates of migration termination (arrival and departure from the breeding or summering ranges). The nesting period was considered to be from 25 May to 5 July.

Breeding behaviour

In geese, it is common for the female to incubate the eggs while the male guards the nest and makes short foraging trips. A pair of geese tracked in 2017 (BFUL059 – a female and BFUL061 – a male) was

studied to confirm these behaviours. During the nesting period, the nest-site of a presumed female was identified by an "asterisk" pattern centred on a specific point less than 100 m away. The presence of a breeding male was inferred if the bird remained in a small territory (1 km in diameter if it was a circular territory and a maximum of 12 km long if it was a riverine territory) for more than 10 days. Data on nest positions were analysed using QuantumGIS 3.16.5, which utilised conic coordinates in the Asia_North_Albers_Equal_Area_Conic projection. In this study, a goose was categorised as a non-breeder if its track did not exhibit any of the aforementioned patterns (male or female type) between 25 May and 5 July. Conversely, if a goose displayed a "nesting" pattern, it was considered to be a breeder, indicating a nesting attempt. The breeding propensity was then calculated as the proportion of geese that attempted to nest out of all the tracked geese. There is a method of estimation of geese nest location via overall dynamic body acceleration (Schreven et al. 2021), however it was not suitable for use in our study due to missing acceleration data.

Nest-sites, nesting habitats, and time of nesting

Each bird was categorised as a breeder or non-breeder (see above). Utilising QuantumGIS 3.16.5, we generated heatmaps from breeding tracks from 25 May to 5 July, and each "bubble" on the heatmap was analysed for dates, bubble size, frequency, and regularity of a bird within a bubble. If a bubble was less than 120 m in diameter and a bird visited it for more than 10 days, the site was deemed a female nest. If a 1 km-long bubble was extended along a river or small stream, it was considered the nest-site of a male.

Nesting habitat was identified using the ESRI Satellite map. Nest-sites were classified into geographical elements such as mountain/hill slopes, river valleys, small stream valleys, islands in the river, lakesides, and so on. To determine the dominant vegetation type of the nesting habitat, the nest positions were plotted on the vegetation map of Russia (Yurkovskaya 2011).

The onset of nest-building and laying was considered to begin on the day when the bird was first recorded inside the nest bubble. For geese that successfully nested (see method below), the last day in the bubble was regarded as the day when the brood left the nest.

Nest success

A bird that stays within a relatively small area and does not exhibit significant flights of over 20 km for a period exceeding 80 days is considered to be a successful breeder. This 80-day threshold includes five days for laying an average clutch of five eggs (Degtyarev and Perfilyev 1996), based on a 30-hour egg-laying interval taken from data for the Greater White-fronted Goose (Ely et al. 2020), 25 days for incubation (Kear 2005), and 46 days for the period from hatching to fledging of the young (Kear 2005). A prenesting period of an unknown duration was not taken into account. An unsuccessful nesting or brood-rearing attempt is determined if the above patterns occur within less than 80 days. Failure of breeding by EPLWFG is indicated by an immediate long flight away from the nest-site; successful breeding geese must stay with the goslings, who are unable to fly (Kölzsch et al. 2019).

Summer timing

For non-breeding geese, it is assumed that the spring migration has transitioned into summer activities when the trans-latitudinal movement shifts to trans-longitudinal. Similarly, summering is considered complete when a prominent southbound movement is detected. For breeding birds, the date of arrival at a breeding site was considered to be the beginning of the summering phase. It is also assumed that the summering phase encompasses breeding, moult, and movements between summering sites.

Summer movements including moult migration

Distances for the summer movements along the route were measured, with local movements at staging sites being disregarded for this specific analysis. To validate the summer site, a four-day rule was consistently applied, which meant that the bird should have spent a minimum of four days at the site for it to be classified as a staging site. A staging site was specifically defined by a bubble with a 50-km diameter. The direct distance between staging sites was then calculated, and the migration distance was defined as the cumulative travel distance.

Moult migration was recognised as a pronounced movement occurring in late June to early July, following which the remigial moult commenced (see section below). Return migration from the moulting site was also recorded for some birds; the return migration was considered if a bird returned to within 100 km of the early summer staging or breeding site after the remigial moult.

Moult timing, duration, and moulting habitats

Moulting sites and the period of flightless moult were identified based on a movement pattern characterised by limited or no flight activity. In this context, if the speed of the bird was less than 15 km/hour, and all hourly movements were less than 2 km, it was indicative of the bird's engagement in moulting activity. The duration of flightless moult was calculated as the period (in days) between the last and the first record of speed over 15 km/hour. Additionally, the time spent on the moulting grounds was measured as the period between the bird's arrival at, and departure from, the moulting site. Moulting habitats are mainly lakes and rivers and were identified by plotting the moulting tracks on a vegetation map (Yurkovskaya 2011). The length of the river stretch occupied by the bird during the entire flightless period was also measured, excluding birds with the Lwfg prefix in 2018.

Site fidelity

Nesting and moulting site fidelity was recorded if the bird returned to within 10×10 km of the same area in the following year/s.

Surveys in Chukotka, Russia for brood size and proportion of juveniles in brood-rearing flocks

During the period of July–August from 2002 to 2021, surveys of rivers were conducted by travelling downstream in a motorboat from the upper reaches of remote rivers, which were accessed by helicopter from the Chaun Biological Station (68.3°N, 170.2°E). A comprehensive description of the study area, survey methods, peak counts for each river, and the proportion of juveniles in brood-rearing flocks during the period 2002–2010 have been published previously (Solovieva and Vartanyan 2011; Tian et al. 2021). Peak counts along nine previously studied rivers ranged from 0.3 to 3.9 individuals/km, as reported by Tian et al. (2021). Brood sizes were recorded in both single broods and brood-rearing flocks (if a family was recognised in a flock). The proportion of young in the EPLWFG

in mid-August, prior to the autumn migration, was calculated based on breeding behaviour, nest success, and brood size obtained from this study (refer to the Results section for further details). The proportion of juveniles in brood-rearing flocks was estimated to be 0.29 in 2002–2010 (Solovieva and Vartanyan 2011).

Remoteness of breeding, moulting, and staging sites

We measured the direct distances from each breeding, moulting, and staging site to the nearest human settlement using the Google Earth ruler.

Results

Summary of tracking results

We received summer tracks from 11 individuals (BFUL prefix) out of the 88 EPLWFG captured in winter 2016/7 (12.5%). The tracking data for the remaining 77 birds were not recovered, either due to device malfunction or casualty. Eleven trackers recorded during the entire summer of 2017. Three trackers worked for an additional year (2018), and one also continued working for a third year (2019). There was a pair of geese (BFUL059 and BFUL061) whose tracks were identical in 2017 except for nesting patterns; we considered this to be one track in the summer of 2017. A total of 15 summer records were available for analysis from the birds captured in winter 2016/7.

Additionally, we received summer tracks from 9 out of the 10 individuals (Lwfg prefix) captured in the summer of 2018 (90%, Table S1; see also Ao et al. 2020). The tracks from these birds were analysed by Ao et al. (2020) for migration chronology and summer staging sites (Table S2). There were nine trackers in the first summer (2018), four trackers in the next summer (2019), and two trackers which continued to work in the third summer (2020).

Thus, we analysed a total of 30 summer tracks from 19 individual EPLWFG, with 9 tracks (Lwfg prefix) representing a part of the moult, post-moulting period, and departure from summer grounds, rather than the full summer. These tracks were not independent as they originated from the same individuals tracked for more than one summer. Some tracks were terminated before the end of the summer (Table S1), so the number of tracks used for each type of analysis differs from the expected 30/19 tracks.

Breeding behaviour

Of the 11 LWFG captured in winter 2016/7 (of unknown age with the BFUL prefix) 36.4% (four birds) were non-breeders during the first year of tracking. Their tracks did not show nest-sites. We assumed that these four non-breeders were AHY/ASY (after hatch year or after second year) birds because they are usually less careful and easier to capture. Among 16 potential breeding attempts by adult geese (a pair is considered as one track), tracked for at least one summer after capture, only one could be classified as a nonbreeder (Lwfg06 in 2020; Table S2). Thus, a breeding propensity of 93.8% was estimated for adult EPLWFG and 76.2% for all geese (16 out of 21) arriving in the summer range.

Nest-sites, nesting habitats, and time of nesting

The tracked EPLWFG nested in various habitats, with most nestsites situated either in the valleys of streams and medium-sized rivers or on mountain slopes facing rivers and streams. Specifically, the nests were placed in stream valleys (six nests), on mountain/hill slopes (three nests), in medium-sized river valleys (three nests), on the shore/island of a lake in a river valley (two nests), or on an ox-bow bank in the river valley (one nest) (Table 1). As a result, 80% of the nests were attributed to the mountainous streams or river valleys, while 20% of nests were situated on hill slopes. Nests were located in various bioclimatic zones, including the northern taiga, sparse forest, sparse forest–tundra, mountain tundra, and flat hypoarctic tundra. Notably, in most of these zones, except for the tundra, the dominant tree species were larches *Larix gmelinii* and *L. cajanderi* (Yurkovskaya 2011) (Table 1). The laying onset was estimated in 12 cases (both successful and failed breeding attempts) and the median date was 30 May \pm 1 day. Broods left two successful nests on 21 June (BFUL062 in 2018) and 27 June (BFUL065 in 2017).

Nest success

Two breeding attempts appeared to be successful: BFUL065 in 2017 and BFUL062 in 2018 seem to have successfully bred and raised their broods to the fledgling stage. The other 13 geese with nesting attempts (both BFUL and Lwfg prefixes) remained at the presumed breeding grounds for an average of 32.2 ± 1.4 days (range 25– 42 days) and then flew off on moult migration, so we assumed that they had failed to nest. Only 13.3% of the 15 tracked birds bred successfully in 2017–2020 (a pair BFUL059 and BFUL061 in 2017 was considered as one breeding attempt). None of the eight geese with the Lwfg prefix, which were brood rearing in summer 2018 (presumably successful breeders), bred successfully in 2019 and 2020 (Table S2).

Summer timing

EPLWFG individuals that attempted to nest arrived at their breeding sites on 23 May \pm 1.7 days on average (range 15 May to 6 June, Table 2; see Table S2 for the start of the breeding period), while nonbreeders delayed their arrival at the summer grounds. For instance, the non-breeding individual BFUL057 moved slowly down the Lena River from 10 June to 26 June, when it reached the Lena Delta. The 26 June was considered to be the date when the spring migration was complete and summering began. An adult individual Lwfg06 was brood rearing (successful breeder) in 2018 and was considered a failed breeder in 2019 (arrival date 26 May). However, this bird arrived on 18 June in 2020, and the track did not show any nesting attempt. As mentioned, this goose may have lost a partner sometime between the breeding seasons of 2019 and 2020. On average, the non-breeding geese reached the first summer stop-over site on 10 June \pm 5.8 days (range 26 May–26 June) (Table 2).

The duration of summering was shortest for non-breeders, averaging 103.0 ± 7.5 days (Table 2). Failed breeders stayed 22 days longer because they arrived earlier in an attempt to nest, and successful breeders stayed even longer (133 days in BFUL062 in 2018; the second track of successful BFLU065 ended on 5 September 2017 before departure), departing nine days later than failed breeders on average. Successful breeders (BFUL065 and Lwfg03-09) departed on 6 October on average (range 28 September-9 October). Four non-breeders departed on 24 September, on average, and birds that failed to breed departed, on average, on 27 September (Table 2). The three-week difference between arrival (*t*-test, df = 4, P = 0.029) and a one-week difference in departure (df = 7, P = 0.004) of breeders and non-breeders were both significant. There was a significant difference between the departure times of successful and unsuccessful breeders, with the latter departing eight days earlier (df = 13, P = 0.004).

 Table 1. Characteristics of breeding sites of the Lesser White-fronted Goose Anser erythropus revealed by tracking in 2017–2020. Coordinates of sites are given in Table S2

ID	Year	Site name	Bioclimatic/vegetation zone	Breeding habitat description	Site fidelity
BFUL050	2017	Buor-Yuryakh R	Larch sparse forest-tundra	mountain slope	NA
BFUL051	2017	Sededema R	a R Larch sparse forest mountain slope		NA
BFUL061	2017	Korkodon R	OBLDPT ¹ river valley		NA
BFUL062	2017	Kychaalaakh R	Larch sparse forest-tundra	stream valley	NA
BFUL065	2017	Khattynaakh R	Larch northern taiga	island on lake in the river valley	NA
BFUL074	2017	Nyronnayvaam R	SFTLDP ²	river valley	NA
BFUL051	2018	Sededema R	Larch sparse forest	stream valley	0.4 km
BFUL061	2018	Korkodon R		mountain slope	4.1 km
BFUL062	2018	Kychaalaakh R	Larch sparse forest-tundra	stream valley	within 4 km
BFUL062	2019	Kychaalaakh R	Larch sparse forest-tundra	stream valley	within 4 km
Lwfg03	2019	Rauchua R	TSF ³	stream valley	Assumed ⁴
Lwfg06	2019	Rauchua R	TSF ³	lakeshore in the river valley	Assumed
Lwfg07	2019	Rauchua R	TSF ³	river valley	Assumed
Lwfg08	2019	Rauchua R	TSF ³	ox-bow bank	Assumed
Lwfg03	2020	Rauchua R	TSF ³	stream valley	1.5 km

¹Okhotsk-Beringian larch– dwarf pine taiga

²East-Siberian sparse forest-tundra with larch and dwarf pine

³East-Siberian tundra with Dryas punctata, D. ajanensis, and steppe fragments

 4 Site fidelity is assumed if the bird returned to within 10 × 10 km of the previous nesting site area in the following year/s.

Table 2. Timing, number of staging sites (average ± SE, *n* in parentheses) and movements of the Eastern Lesser White-fronted Geese Anser erythropus on the breeding grounds, 2017–2020

Parameter	Non-breeder	Unsuccessful breeder	Successful breeder
Arrival summer range ¹ (Julian date)	160.4 ± 5.8 (5)	142.3 ± 1.7 (13)
Departure from summer range (Julian date)	267.5 ± 3.0 (5)	269.1 ± 1.7 (15)	278.5 ± 1.4 (8)
Duration of summering, days	103.8 ± 7.5 (4)	125.5 ± 2.7 (11)	133 (1)
Number of summer staging sites ²	3.4 ± 0.9 (5)	3.9 ± 0.2 (12)	2 ± 0.0 (8)
Distance travelled during summer along the route (km)	1253 ± 248 (5)	1281 ± 176 (12)	79 ± 13 (8)
Average speed of summer movement (km/day)	203 ± 95 (4)	265 ± 42 (8)	140; 24 (2) ³

¹All breeders combined (successful and unsuccessful)

²Including breeding and moulting sites:

³Known for BFUL062 in 2017 and BFUL062 in 2018.

Summer movements including moult migration

Non-breeding EPLWFG fly within the summering range until they stop to moult (Figure 1). It can be challenging to detect staging sites during their uninterrupted summer movement. Applying the four-day rule, we can conclude that non-breeders used an average of 3.4 ± 0.9 sites per summer and travelled an average of $1,253 \pm 248$ km (range 544-1,781 km) (Table 1 and Figure 1). Failed breeders used an average of 3.8 ± 0.2 sites per summer, including breeding and moulting sites, and flew an average of $1,281 \pm 176$ km during the summer (range 296-2,127 km) (see Table 1 and Table S2). Non-breeders and failed breeders travelled similar distances during the summer (the difference was insignificant, P = 0.35). The successful breeders made a short move to another site (average 74.4 ± 14.7 km for all eight successful breeders) after their broods and themselves gained flight ability in mid-August (Table 1). Their autumn staging sites were situated north of the breeding sites, and the duration of stay at these sites averaged 47 ± 2 days (range 35–53 days; n = 8).

Twelve geese left their nest-sites after nesting or brood failure and began the moulting migration, covering an average distance of 751 ± 114 km in 3.9 ± 0.8 days (range 186-1,330 km, 0.5-8.5 days) (Table S2). For two non-breeders, moult migration was part of their continuous, non-stop flight, while the other two flew distances of 475 and 1,245 km to reach their moulting grounds (Figure 1). An adult non-breeder (lwfg06 in 2020) flew 580 km from Rauchua River (breeding site in previous years) to the moulting site. All birds migrated to the north or north-west from their breeding or staging sites to the moulting grounds.

Out of 16 tracks of failed breeders on the moult migration, seven (43.7%) tracks indicated a return migration to their early summer staging site after completing the moult (Table S2). The distance of



Figure 1. Individual routes of 11 LWFG tracked from wintering grounds (prefix BFUL) in the summering range in 2017–2019 (A) and summer tracks of 8 LWFG tracked from summer grounds (prefix Lwfg) in 2018–2020 (capturing site is shown by open triangle; B). Summer tracks were considered after the bird crossed 600N during spring migration and before it crossed 600N during autumn migration. Each bird is denoted by a unique color.

the return migration was equal to that of the moult migration. The return migration took an average of 7 ± 3 days (range 2–29 days; n = 7) and included a staging on the way for three out of four geese. On average, the return migration was slower than the moult migration (average 4 ± 1 days) of the same individuals in the same years.

Moult timing, duration, and moulting habitats

The EPLWFG that successfully breed moulted near their nest-sites and formed brood-rearing flocks of up to 70 individuals, often mixed with Bean Geese *Anser fabalis* (our observations in West Chukotka). After hatching, successful breeders do not fly long distances and take flight alongside their young, making the duration of their flightless moult unclear. In this study, successful breeders typically began flying on the median date of 13 August (range 8–16 August; n = 9; Table 3). In West Chukotka, the earliest young and brood-rearing adults were observed to fly between 6 August and 12 August (2010–2019).

Out of 20 tracks of non- and failed breeders (excluding bird Lwfg03 in 2018 from Rauchua River), 14 cases (70%) of moult were observed along the lower reaches of the San-Yuryakh and Kyuanekhtyakh Rivers at 72.3°N, and three cases (15%) of moult were observed along the Malaya Kuropatoch'ya River at 70.8°N (Table 3

Table 3. Flightless moult timing and habitats of non- (NB) and failure breeding (FB) Lesser White-fronted Geese *Anser erythropus* in 2017–2020. Coordinates of moulting sites are given in Supplementary material Table S2 (site status = mo). The portion of river used for moulting by individual geese was measured along the river ignoring visits to nearby lakes

Bird ID	Status	Moult start date	Moult end date	Flightless duration (days)	Moulting habitat	Location	Portion of river (km)
BFUL041	NB	02.07.2017	06.08.2017	35	river	Kyuanekhtyakh R	26.7
BFUL044	NB	28.06.2017	03.08.2017	36	river	San-Yuryakh R	24.6
BFUL050	FB	16.07.2017	14.08.2017	29	river	Muksunokha R	12.7
BFUL051	FB	28.06.2017	30.07.2017	32	river	Kyuanekhtyakh R	29.1
BFUL057	NB	12.07.2017	14.08.2017	33	swamp	Lena Delta	12.2
BFUL059	FB	30.06.2017	09.08.2017	40	river	San-Yuryakh R	26.4
BFUL061	FB	30.06.2017	01.08.2017	32	river	Kyuanekhtyakh R	26.4
BFUL062	FB	03.07.2017	30.07.2017	27	river	San-Yuryakh R	9.3
BFUL068	NB	03.07.2017	17.08.2017	45	river	Kyuanekhtyakh R	32.3
BFUL074	FB	28.06.2017	06.08.2017	39	river	Kyuanekhtyakh R	35.4
BFUL051	FB	30.06.2018	24.07.2018	24	river	Kyuanekhtyakh R	19.7
BFUL061	FB	06.07.2018	09.08.2018	34	river	Korkodon R	4.8
BFUL062	FB	30.06.2019	31.07.2019	31	river	San-Yuryakh R	8.6
Average ± SE		30 June ± 1.5	6 August ± 2.0	31.6 ± 1.6			20.6 ± 2.8
Lwfg03	FB	NA	15.8.2018	NA	river	Rauchua R	NA
Lwfg03	FB	9.7.2019	3.8.2019	26	lake	San-Yuryakh R	NA
Lwfg03	FB	7.7.2020	28.7.2020	21	river	San-Yuryakh R	19.7
Lwfg06	FB	14.7.2019	3.8.2019	20	river	M. Kuropatoch'ya R	13.9
Lwfg06	NB	5.7.2020	11.8.2020	38	river	M. Kuropatoch'ya R	30.3
Lwfg07	FB	25.6.2019	16.7.2019	21	lake	M. Kuropatoch'ya R	NA
Lwfg08	FB	2.7.2019	25.7.2019	23	river	San-Yuryakh R	22.4
Average ± SE				24.8 ± 2.8			21.6 ± 3.4

and Table S2; see also Lwfg01 in Ao et al. 2020). These key moulting sites are situated in the northern hypoarctic tundra zone and are characteristic of lowland polygonal marshes (Yurkovskaya 2011). Non- and failed LWFG with the BFUL prefix were flightless for an average of 31.6 ± 1.6 days (range 24–40 days; Table 3), while geese with the LWFG prefix were flightless for an average of 24.8 ± 2.8 days (range 21–38 days). The difference between the groups is statistically significant (*t*-test; *P* = 0.012). The earliest a bird (regardless of prefix) became flightless was on 25 June, and the latest a bird finished moult was on 17 August. EPLWFG utilised an average of 20.8 ± 2.9 km of river during the moult (Table 3).

Site fidelity

We utilised data from four EPLWFG (BFUL051, BFUL061, BFUL062, and Lwfg003) tracked for more than one season for estimating nest-site fidelity. All four returned to their previous breeding area in subsequent year/s. Distances between consecutive nests of the same bird were estimated in three cases, averaging 2 km (Table 1). BFUL062 (presumably male), tracked for three years, nested within 4 km along a small stream (the exact nest position is unclear) in all three summers. This bird was successful (fledgling young) only once, in 2018; breeding failed in 2017 and 2019. All five adult geese captured in the summer of 2018 on the Rauchua River also made nesting attempts there in 2019 and 2020. Lwfg06, which

did not breed in 2020, also visited the Rauchua River briefly. Thus, we estimate that breeding-site fidelity is 100% in adult LWFG and does not seem to depend on the breeding success of the previous season.

Five out of six (83.3%) of failed breeding adult LWFG also demonstrated fidelity to their moulting site. One goose (BFUL061 in 2018) seemed to have missed the time to reach the moulting site on the Kyuanekhtyakh River (used in 2017) after losing a brood on 3 July 2018 and becoming flightless on 6 July (Table 3). If we exclude the case of BFUL061, fidelity to the moulting site could be estimated to be 100%.

Brood size and proportion of juveniles in brood-rearing flocks

According to our observations in West Chukotka, the LWFG brood averaged 3.69 ± 0.19 young, aged 5–7 weeks old, in 2003–2021 (range 2–7; n = 45). The proportion of juvenile birds was 46.9% of all the birds reported during the surveys at brood-rearing habitats (medium-sized rivers of the West Chukotka). Recently, the annual productivity of EPLWFG can be estimated at 0.49 young per adult pair. We estimated the proportion of young in the EPLWFG population in mid-August (prior to fledging) as 16.5%. This was determined by halving the overall pair productivity of 0.49, which was calculated from multiplication of the breeding rate (76.2%), nest success (13.3%), and number of young per successful brood (3.69). Brood survival from hatch to 5–7 weeks old is unknown and thus it is not considered in the estimation of productivity.

Remoteness of breeding, moulting, and staging sites

There is no difference in remoteness (P = 0.17) between the Rauchua breeding site, where EPLWFG returned to nest in the years following the capture year, and other breeding sites, discovered by tracking, The average distance of the summer staging sites from the nearest village was 69.7 ± 4.8 km (n = 46). The remoteness of the breeding sites was higher, averaging 144 ± 1.8 km (n = 16), and the moulting sites were the most remote, situating on average 159 ± 14.5 km (n = 20) from the nearest village. There was a significant difference among the groups (One-way ANOVA, F _{2.78} = 27.2; P < 0.001).

Discussion

This paper presents new data on the nesting and moulting ecology of the relatively unknown EPLWFG obtained from GPS/GSM tracking and field surveys. It provides estimates of breeding propensity, nest success, productivity, and nesting and moulting site fidelity for the first time. Tracking the geese led to the discovery of remote moulting sites important for most of the population, which were previously unknown. Literature data on nesting habitats and the summer timing of EPLWFG were scarce and this paper fills the gap in knowledge.

Breeding propensity

Boom et al. (2023) found that tracking was an effective way to document breeding behaviour of geese, which is also true for the EPLWFG. Breeding propensity is one of the most mysterious demographic parameters and is almost impossible to quantify in wild birds, even in marked populations (Jean-Gagnon et al. 2018). This parameter could be estimated as the proportion of birds that commenced nesting upon reaching the breeding grounds. Among the 16 potential breeding attempts by adult EPLWFG, only one failed, presumably after losing its partner. This individual arrived at the nesting site late on 18 June, compared with 26 May in the previous year (see Lwfg06 in 2020 in Table S2). Breeding propensity in adult geese in the Arctic varies from 22% to 100%, primarily depending on spring weather conditions (Boom et al. 2023; Reed et al. 2004; Sedinger et al. 2008). Geese in temperate latitudes are less dependent on spring weather for their breeding intention (Boom et al. 2023). The EPLWFG breeds in the sub-Arctic and boreal forest where spring conditions seem to be favourable for the annual nesting effort of all adult individuals. However, the small number of tracked individuals in this study does not allow for accurate estimates of breeding rates. Any way we estimate breeding propensity in the EPLWFG in 93.8% in adults and 76.2% in all tracked birds.

Nest-sites, nesting habitats, and time of nesting

All geese (LWFG prefix) captured as brood-rearing adults in summer on the Rauchua River (tundra biome) returned to the same site in the following year/s, so we did not analyse their nesting habitats. All seven EPLWFG tracked from wintering grounds (BFUL prefix and including Lwfg01 from Ao et al. 2020) nested in taiga or forest– tundra biomes (sometimes in intrazonal mountain tundra), and none of them nested in the tundra biome (Table 1). This pattern is consistent with previously known nesting or brood-rearing records from taiga and forest-tundra on the Omolon and Muna Rivers (Andreev 2001; Egorov and Okhlopkov 2007) and the "forest edge" in Yakutia (see reviews in Degtyarev and Perfilyev 1996; Morozov and Syroechkovski 2002). The breeding population of the Chaun-Rauchua Lowland, of the northern tributaries of the Anadyr River (Tian et al. 2021), and a single record of a brood near Sellyaskaya Bay of the Laptev Sea (Bysykatova and Krapu 2009) were the only breeding EPLWFG recorded from the tundra zone. Low altitude mountains, river valleys, and small stream valleys seem to be the major nesting habitats (Morozov and Syroechkovski 2002) (Table 1). It is unclear if the EPLWFG is associated with Peregrine Falcon in nesting, as was proposed for the Western Main population (Korobitsyn and Tyutenkov 2023; Pokrovskaya et al. 2023). Nesting Peregrine Falcons are common on cliffs along rivers with broodrearing EPLWFG in West Chukotka (based on our data).

The median date of commencement of egg-laying is 30 May, which is consistent with egg-laying on 20–25 May for the first pairs in Norway (Øien and Aarvak 2008) and slightly earlier than, on average, 4 June (21 May–13 June) in Finland (Markkola and Karvonen 2020). However, our estimate of egg-laying onset was obtained via spatial–temporal analyses of tracks, in which the first visit to the nest-site may have happened earlier than laying onset, thus our estimate may be biased (earlier) by a few days. More tracking of the EPLWFG is required for modelling the climatic predictors of timing of breeding. Finnish LWFG were found to start egg-laying independently of local weather conditions (Markkola and Karvonen 2020).

Nest success

The 13.3% proportion of successful breeders was lower than the long-term average of 46% in the Fennoscandian population (Aarvak and Øien 2018). In Fennoscandia there was culling of red fox Vulpes vulpes, wolverine Gulo gulo, and lynx Felis lynx, and brown bear Ursus arctos and wolf Canis lupus were absent. All the abovementioned predators, with Arctic fox V. lagopus and sable Martes zibellina, are common, and sometimes numerous, in the breeding range of the EPLWFG. Lower nest success would appear to result from high predation rates from mammalian and avian predators, however documentation of nest predation is absent for the EPLWFG. Nest success in Finland was highly dependent on cumulative positive temperature and cold spells during incubation, with the latter acting negatively on reproductive success (Markkola and Karvonen 2020). The entire breeding range of the EPLWFG is situated in a colder climatic zone than Finland, and there are greater chances of low temperatures in June and cold spells on any day in (https://www.climate-charts.com/World-Climatesummer Maps.html#temperature). This, in part, may explain the low nest success obtained in this study. Tracking of additional individuals is required to study climatic effect on nest success.

Summer timing

An average breeding EPLWFG arrives at its breeding site at 64– 70.5°N on average 17 days earlier than the Greater White-fronted Goose nesting at a similar longitude but at 71–73°N (Table 2); Greater White-fronted Goose arrive on average 162 ± 10 Julian date (Deng et al. 2019). The summering duration in breeding of LWFG is 23–30 days and longer than that of Greater White-fronted Goose, with the departure of LWFG delayed by 6–13 days (for both failed and successful breeders). This is probably due to the more southerly breeding sites of LWFG, allowing for an earlier arrival and longer stay than the Arctic-nesting Greater White-fronted Goose.

LWFG assumed to be younger than two years old and not participating in reproduction arrive later and depart earlier than the adult birds, distinguished by their breeding attempts (Table 2). The three-week difference in arrival and one week difference in departure of breeders and non-breeders were both significant in this study. We lack the tracking data for the same individual as a nonbreeder and then becoming a breeder in the following year, while we may get some insight from an individual tracked from the wintering ground in Poyang Lake, China (Lwfg01 in Ao et al. 2020; not included in the analyses in this study due to lack of permission). It arrived at the summer grounds on 30 June in the first year when it was presumably a sub-adult bird; it arrived at the summer grounds on 25 May in the following year when it had reached reproductive age, presumably two years old (see Cramp and Simmons 1977 for the age of first reproduction). There was a significant difference between the departure times of successful and unsuccessful breeders, with the latter departing eight days earlier. However, this difference might be biased due to seasonal differences in autumn conditions. In two individual birds tracked for more than one summer (Lwfg006 and Lwfg007), the departure after unsuccessful breeding (26 September 2019) was earlier than that after successful breeding (9 October 2018). However, by contrast, a third individual (Lwfg002 failed to breed in both years) departed on 8 October 2018 and 27 September 2019. The later departure in 2018 is more likely a result of a warmer autumn compared with that of 2019. Following the "frost wave hypothesis" (Xu and Si 2019), we suggest that the LWFG leaves summer areas after the onset of frost, which varies between years.

Summer movements including moult migration

The EPLWFG are highly mobile during the summer, using an average of 2 sites for successful breeders, 3.4 sites for failed breeders, and 3.9 sites for non-breeders (see Table 2). Their mobility makes conservation efforts on summer grounds less effective than one would expect unless a network of seasonally protected areas is developed. However, prior to the development of modern bird-tracking techniques, their movement patterns were unclear (Figure 1). Early reconstructions of the summer range based on records of adult birds, nests, and broods were incomplete and suggested a patchy range (Morozov and Syroechkovski 2002), whereas the track-based model suggests the possibility of a continuous summer range (Tian et al. 2021). Typically, in non-breeders (mostly birds younger than two years and 6.2% of adult birds), spring migration is terminated when they begin their remigial moult.

Failed breeders did not stay close to their nests and began the long-distance migration to moult sites. Among 13 failed breeders, only one moved a short distance from the nest-site to moult in the same river catchment. This short moult migration may be attributed to the late loss of a nest/brood and a lack of time to reach more distant moulting sites. Summer staging sites are widely dispersed within the summer range, sometimes in proximity to breeding or moulting sites, although not always (Table S2).

Many failed breeding EPLWFG (43.7%; Table S2) migrated back to their early summer breeding/staging site after completing the moult, thus extending their stay in the summer range. Of the five individuals tracked for more than one summer, three, being failed breeders, never returned to the breeding site after moulting, while two always did. It is difficult to explain why they almost precisely return to the breeding site, sometimes covering over 1,000 km on the return migration (Figure 1 and Table S2). One possible explanation could be fidelity to the autumn migration route, where geese are familiar with stop-over sites; however, the consistency of migration routes among years has not been analysed. Another hypothesis suggests checking conditions for the next breeding season, similar to the Arctic Rough-legged Buzzard *Buteo lagopus* (Curk et al. 2022), but the kind of conditions geese are able to evaluate (food availability, habitat status, hydrology, predators) remains unexplained.

Moult timing, duration, and moulting habitats

The core moulting site for the entire EPLWFG was identified through tracking (Ao et al. 2020; this study). It is situated along the lower reaches of the San-Yuryakh and Kyuanekhtyakh Rivers near their joint estuary, where they flow into Omulyakhskaya Bay of the East-Siberian Sea. The site spans an area of 647 km², approximately 44×16 km, and is considered the "absolute pole of inaccessibility" even within the remote and sparsely inhabited Arctic Siberia (P. Nikol'skiy, personal communication). Located 150-200 km from Yukagir village, reaching this area in summer is only possible using expensive transportation means, such as large ATVs or helicopters. It is not surprising then that 70% of non- and failed breeding EPLWFG gather in this area. All EPLWFG migrated 185-1,330 km to reach optimal moulting habitats, indicating that moult migration is typical for this species. An even longer moult migration of at least 1,800 km was reported for Fennoscandian LWFG (the individual PTT24676 was tracked from Norway to the Taimyr Peninsula; Aarvak and Øien 2003). Another moulting site, the Malaya Kuropatoch'ya River (a small river 150 km long, flowing into the East-Siberian Sea), was utilised by 15% of failed breeders, specifically by two adults (one of them, twice) known to breed 600 km apart on the Rauchua River. This site is located 272-322 km from the nearest village and is almost inaccessible during the summer, akin to the San-Yuryakh and Kyuanekhtyakh Rivers. In contrast to the breeding habitats, the moulting sites are situated in the tundra zone near the Arctic coast (Yurkovskava 2011).

In their review of Anatidae-moulting strategies, Fox et al. (2014) proposed that geese moult in the Arctic to avoid the predators that are abundant at lower latitudes. We would like to emphasise that the moulting sites of the EPLWFG provide both potentially lower predation pressure compared with breeding areas and a lack of human disturbance. The EPLWFG prefer to moult on rivers (17 out of 20 birds; the rest moulted on lakes and swamps; Table 3), likely due to the same safety reasons. We cautiously suggest that the portion of the river they use for moulting, on average 20.8 \pm 2.9 km, seemingly satisfies their food requirements. Although the food and foraging ecology of moulting EPLWFG is unstudied, it could possibly be obtained from further analyses of existing tracks.

Moult duration in the LWFG is not well-documented, although Aarvak and Øien (2003) reported that a male from the Fennoscandian population remained at its moulting site on Kolguyev Island for 33 days in 1997. The significant difference between the flightless period in geese with BFUL and Lwfg prefixes is due to differences in the fix interval. The two-hour interval between fixes in BFUL geese does not allow for the detection of first flight activity, as is possible with 10-minute interval fixes. This first flight after the flightless moult may be just a short test of flying ability and might not be related to movement from the moulting site. Field observations suggest that birds stay flightless after completing the remigial moult in order to accompany their late moulting conspecifics. During the flightless period, undisturbed flocks of LWFG may not fly until the last individual has completed its moult. While the nesting biology of geese could be better understood by using trackers with hourly or higher intervals between fixes, a 10-minute fix interval is preferred for details of the remigial moult. The flightless period of the EPLWFG obtained from trackers with a 10-minute interval is approximately 24.8 \pm 2.8 days, which is close to the 21 days needed for waterfowl birds to replace their flight feathers (Hohman et al. 1992). The moulting window for non-breeding and failed geese lasts from 25 June to 17 August. Successful breeders typically complete their moult around the median date of 13 August, similar to latemoulting failed breeders.

Site fidelity

Adult LWFG showed 100% fidelity to their nesting site despite a previous year nest failure: no evidence of nesting area change was observed during this study. The site fidelity of unsuccessful breeders to the moulting site was also 100%, based on tracking five adult geese for more than one summer. We propose that the summer range of the LWFG includes breeding enclaves comprising hundreds of pairs (known enclaves are the Rauchua River; Solovieva and Vartanyan 2011 and the Muna River; Egorov and Okhlopkov 2007) surrounded by areas with low breeding densities (single broods or small brood-rearing flocks reported during surveys; see also Tian et al. 2021), and concentrated moulting sites. Intriguingly, none of the geese tracked from the wintering grounds nested in the two known breeding enclaves, hinting at the existence of additional enclaves. The tracking results from the wintering grounds suggest the existence of one more breeding enclave on the Sededema River, East Yakutia (BFUL051 in 2017 and 2018 and Lwfg01 in Ao et al. 2020). Given the observed 100% site fidelity to nesting sites in the EPLWFG, identifying additional breeding enclaves should be a conservation priority in the near future.

Brood size and proportion of juveniles in brood-rearing flocks

From 2002 to 2021, the average brood size near the fledging stage was 3.69 ± 0.19 goslings in West Chukotka. This surpasses the average of 3.0 goslings recorded from Norway before red fox control (Øien and Aarvak 2009), the 2.93 goslings in Finnish Lapland from 1989 to 1996 (Markkola and Karvonen 2020), and the 2.86 goslings in the reintroduced Swedish population from 1999 to 2003 (Andersson 2004). The EPLWFG appears to be more productive in brood size than the Fennoscandian sub-population. Additionally, Degtyarev and Perfilvev (1996) observed that a brood averaged 4.2 goslings between hatching and fledging in the Khroma-Indigirka tundra from 1960 to 1962. This brood size is similar to that of the Western Main sub-population, with an average brood of 4.0 goslings (Romanov 2004). In recent findings, the annual productivity of EPLWFG can be estimated at 0.49 young/nesting pair or 16.5 % of young in mid-August. Wang et al. (2021) reported that the proportion of young (after hatch year) EPLWFG in wintering flocks was 15.2 ± 11.2%. These proportions align, assuming a low, seemingly unknown mortality of young during the first migration.

Further evidence of breeding enclaves is the observed proportion of juveniles on the rivers of West Chukotka at 46.9%, which is much higher than the estimated 16.5% of young in the entire EPLWFG. In late summer, adult brood-rearing geese and their young are well separated from the rest of the adult geese who departed from the breeding rivers for moulting.

Remoteness of breeding, moulting, and staging sites

There is no difference in remoteness (P = 0.17) between the Rauchua breeding site, where Lwfg geese returned to nest in the years following the capture year, and other breeding sites discovered by tracking. During ground surveys, all flightless groups of the EPLWFG (broods and moulting non-breeders) were located more than 50 km away from villages or mines in West Chukotka. This distance is significant as local people typically travel along the rivers and seacoast in a one-day motorboat trip to fish, hunt reindeer, and pick mushrooms and berries. Further evidence comes from the largest aggregation of mainly brood-rearing birds, with a peak count of 416 geese in 2018, which occurred along the uninhabited Rauchua River 15 years after the closure of the gold mine and settlement of Baranikha (Solovieva and Vartanyan 2011), a finding confirmed by this study. When the mine and village were active, no LWFG were found along the river (see details of the 1983 survey in Krechmar et al. 1991).

Egorov and Okhlopkov (2007) reported a breeding population on the uninhabited Muna River in Central Yakutia and noted the absence of LWFG along the nearby Molodo River, which has similar habitats but where intensive geological surveys were in progress. The absence of people has been suggested as a key factor in the stability of the breeding population of LWFG in the Putorana Plateau (Western Main sub-population; Romanov 2003).

When selecting staging sites, flying birds appear to be less concerned about human presence: the distance of a staging site to the nearest village is significantly shorter than that of breeding or moulting sites. Even though the EPLWFG summers in the least populated provinces of Siberia (i.e. Yakutia, Chukotka, and Magadan Oblast) with an average density of 0.23 people/km, they actively select the most remote areas for nesting, especially for remigial moult. The model of EPLWFG summer range shows that human presence affects site suitability, with a decrease in species presence starting around 160 km from human settlements (Tian et al. 2021). Locating sites situated 160 km or more away from any settlement is not an easy task, even in unpopulated East Siberia, thus the longdistance summer travels of the EPLWFG are not surprising.

Conservation recommendations

- 1. Strong nest-site fidelity of adult birds contributes to the formation of local breeding populations, which might be conservation units if genetic studies support this differentiation.
- 2. The EPLWFG selects the remotest and least human-accessible area for the remigial moult, and the key site was discovered only with the help of remote sensing. To enhance conservation management, we suggest that no ground-based expedition to this area should be undertaken during moulting periods (July–August), although aerial surveys to count birds are possible.
- 3. As summering sites are well protected by their extreme remoteness, conservation efforts are best concentrated on the wintering areas and migration stop-over sites.
- 4. Appropriate high-resolution GPS/GPRS tracking does not seems to reduce breeding behaviour in the EPLWFG and is recommended as the best method for investigating migratory ecology, breeding and moulting habitats, breeding propemsity, and nest success in the remote areas of the summer range. However, extreme care should be taken when tracking rare species such as the LWFG.

Supplementary material. The supplementary material for this article can be found at http://doi.org/10.1017/S0959270924000285.

Acknowledgements. We gratefully acknowledge the contribution of the East Dongting Lake Nature Reserve and the hard work of the field teams in China and Russia, Carlie Strancki and Thomas Ogilvi. Logistic support and transportation to remote rivers of the West Chukotka was provided by Chukotka Gold Mining Co., the subsidiary of Kinross Gold, with personal acknowledgement to Claude Schimper, Evgenia Saevich, and Daria Krivolapova. The following information was supplied relating to ethical approvals (i.e. approving body and any reference numbers): approval by capture of adult birds was conducted with permission from the Ministry of Natural Resources of Russian Federation (#058 from 10 May 2018); tracking of birds was permitted by the Federal Service for Technical and Export Control (Permit No 320). Approval for bird capture and transmitter deployment was obtained from the Forestry Department of Hunan Province, China, under scientific research licence (No.11 Xiang Forest Protection 2014). Field research was conducted with permission from the Bureau of East Dongting National Nature Reserve. Track maps for the birds with the prefix lwfg and tracking information for the birds with the prefix BFUL can be submitted for review but cannot be published due to concern for the safety of endangered birds. Our study was supported by the National Key Research and Development Program of China (2017YFC0405300) and the National Natural Science Foundation of China (Grants No. 31870369). Tracking from breeding grounds (lwfg prefix) was supported via a joint project of DS and FM, State Key Laboratory of Urban and Regional Ecology, Research Center for Eco-Environmental Sciences, Chinese Academy of Sciences. The funders had no role in the study design, data collection and analysis, decision to publish or preparation of the manuscript. Authors' contributions: DS and JL contributed equally to this work, and carried out the fieldwork, analysed the data, wrote the paper, and prepared the figures and tables. GD and SV carried out the fieldwork. DB prepared the figures and revised the paper. HT, RF, FM, and CL revised the paper. QZ and GL conceived and designed the experiments, contributed materials/analysis tools, and revised the paper. All authors contributed critically to the drafts and gave final approval for publication.

References

- Aarvak T. and Øien I.J. (2003). Moult and autumn migration of non-breeding Fennoscandian Lesser White-fronted Geese Anser erythropus mapped by satellite telemetry. Bird Conservation International. 13, 213–226. https:// doi.org/10.1017/S0959270903003174
- Aarvak T. and Øien I.J. (2018). D2 Lesser White-fronted Goose Anser erythropus – Fennoscandian population. In Fox A.D. and Leafloor J.O. (eds), A Global Audit of the Status and Trends of Arctic and Northern Hemisphere Goose Populations. Akureyri: Conservation of Arctic Flora and Fauna International Secretariat, pp. 40–42.
- Andersson A. (2004). The reintroduction of the Lesser White-fronted Goose in Swedish Lapland – a summary for 2000–2003. In Aarvak T. and Timonen S. (eds), *Fennoscandian Lesser White-fronted Goose Conservation Project. Report 2001–2003.* WWF Finland Report No 20 & Norwegian Ornithological Society. NOF Rapportserie Report No 1-2004. Trondheim: Birdlife Norway, pp. 51–52.
- Andersson Å. and Holmqvist N. (2010). The Swedish population of Lesser White-fronted Goose Anser erythropus – supplemented or re-introduced. Ornis Svecica 20, 202–206. https://doi.org/10.34080/os.v20.22620
- Andreev A.V. (2001). Wetlands in Northeastern Russia. Wetlands in Russia, vol. 4. Moscow: Wetlands International.
- Ao P., Wang X., Solovyeva, D., Meng F., Ikeuchi T., Shimada T. et al. (2020). Rapid decline of the geographically restricted and globally threatened Eastern Palearctic Lesser White-fronted Goose *Anser erythropus*. *Wildfowl* SI6, 206–243.
- Artiukhov A.I. and Syroechkovski Jr E.E. (1999). New data on distribution of Lesser white-fronted goose in the Abyi Lowland (Eastern Yakutia). *Casarca* 5, 136–143.
- BirdLife International.(2018). *Anser erythropus*. The IUCN Red List of Threatened Species 2018: e.T22679886A132300164. Available at https://www.iucn redlist.org/species/22679886/132300164 (accessed 15 November 2021).
- Boom M.P., Schreven K.H.T., Buitendijk N.H., Moonen S., Nolet B.A., Eichhorn G. et al. (2023). Earlier springs increase goose breeding propensity and

nesting success at Arctic but not at temperate latitudes. *Journal of Animal Ecology* **92**, 2399–2411. https://doi.org/10.1111/1365-2656.14020

- Bysykatova I.P. and Krapu G.L. (2009). Breeding conditions report for Mainland east of Yana River Delta, Yakutia, Russia, 2009. In Soloviev M. and Tomkovich P. (eds), Arctic Birds: An International Breeding Conditions Survey. Available at http://www.arcticbirds.net/info09/ru270ru46509.html (accessed 25 October 2022).
- BirdLife International (2021). Anser erythropus (Europe assessment). The IUCN Red List of Threatened Species 2021: e.T22679886A166193479. https://dx.doi.org/10.2305/IUCN.UK.2021-3.RLTS.T22679886A166193479. en. Accessed on 19 December 2024.
- Cao L., Fox A.D., Morozov V., Syroechkovskiy E. and Solovieva D. (2018). D1 Eastern Palearctic Lesser White-fronted Goose Anser erythropus. In Fox A.D. and Leafloor J.O. (eds), A Global Audit of the Status and Trends of Arctic and Northern Hemisphere Goose Populations. Akureyri: Conservation of Arctic Flora and Fauna International Secretariat, pp. 38–39.
- Cramp S. and Simmons K.E.L. (1977). *The Birds of the Western Palearctic*, vol. 1. Oxford: Oxford University Press.
- Curk T., Kulikova O., Fufachev I., Wikelski M., Safi K. and Pokrovsky I. (2022). Arctic migratory raptor selects nesting area during the previous breeding season. *Frontiers in Ecology and Evolution* **10**, 865482. https://doi.org/10. 3389/fevo.2022.865482
- Degtyarev A.G. and Perfilyev V.I. (1996). The Lesser White-fronted Goose (*Anser erythropus*) in Yakutia. *Casarca* 2, 113–124. (in Russian with English summary)
- Deng X., Zhao Q., Fang L., Xu Z., Wang X., He H. et al. (2019). Spring migration duration exceeds that of autumn migration in Far East Asian Greater Whitefronted Geese (*Anser albifrons*). Avian Research 10, 19. https://doi.org/10. 1186/s40657-019-0157-6
- Egorov N. and Okhlopkov I. (2007). New data on nesting of the white-fronted goose (*Anser erythropus*) from Yakutia. *Zoologichesky Zhurnal* **86**, 1482–1485 (In Russian with English summary)
- Ely C.R., Dzubin A.X., Carboneras C., Kirwan G.M. and Garcia E.F.J. (2020). Greater White-fronted Goose (*Anser albifrons*), version 1.0. In Billerman S. M. (ed.), *Birds of the World*. Ithaca: Cornell Lab of Ornithology. https://doi. org/10.2173/bow.gwfgoo.01
- Fox A.D., Flint P.L., Hohman W.L. and Savard J-P.L. (2014). Waterfowl habitat use and selection during the remigial moult period in the northern hemisphere. *Wildfowl* SI4, 131–168.
- Fox A.D. and Leafloor J.O. (eds) (2018). A Global Audit of the Status and Trends of Arctic and Northern Hemisphere Goose Populations. Akureyri: Conservation of Arctic Flora and Fauna International Secretariat. Available at https:// pubs.usgs.gov/publication/70196227.
- Hohman W.L., Ankney C.D. and Gordon D.H. (1992). Ecology and management of postbreeding waterfowl. In Batt B.D.J., Afton A.D., Anderson M.G., Ankney C.D., Johnson D.H., Kadlec J.A. and Krapu G.L. (eds), *Ecology and Management of Breeding Waterfowl*. Minneapolis: University of Minnesota Press, pp. 128–189.
- Ikawa M.J. and Ikawa H. (2009). Lesser White-fronted Goose Anser erythropus at Sarobetsu in northern Hokkaido, Japan: a preliminary report on numbers in autumn. Ornithological Science 8, 131–138. https://doi.org/10.2326/ osj.8.131
- Jean-Gagnon F., Legagneux P., Gilchrist G., Bélanger S., Love O.P. and Bêty J. (2018). The impact of sea ice conditions on breeding decisions is modulated by body condition in an arctic partial capital breeder. *Oecologia* 186, 1–10. https://doi.org/10.1007/s00442-017-4002-5
- Kear J. (ed.). (2005) Ducks, Geese and Swans, vol. 1: General Chapters, and Species Accounts (Anhima to Salvadorina). Oxford: Oxford University Press.
- Kölzsch A., Müskens G.J.D. M., Szinai P., Moonen S., Glazov P., Kruckenberg H. et al. (2019). Flyway connectivity and exchange primarily driven by moult migration in geese. *Movement Ecology* 7, 1–11. https://doi.org/10.1186/ s40462-019-0148-6
- Korobitsyn I. and Tyutenkov O. (2023). New data on the distribution of Redbreasted goose and Lesser white-fronted goose in the south of the Gydan Peninsula. Acta Biologica Sibirica 9, 379–386. https://doi.org/10.5281/ zenodo.7927528
- Krechmar A.V., Andreev A.V. and Kondratiev A.V. (1991). Birds of the Northern Plains. Moscow: Nauka. (In Russian, English summary)

- Lei J., Jia Y., Wang Y., Lei G., Lu C., Saintilan N. et al. (2019a). Behavioural plasticity and trophic niche shift: how wintering geese respond to habitat alteration. *Freshwater Biology* **64**, 1183–1195. https://doi.org/10.1111/fwb.13294
- Lei J., Jia Y., Zuo A., Zeng Q., Shi L., Zhou Y. et al. (2019b). Bird satellite tracking revealed critical protection gaps in East Asian–Australasian Flyway. *International Journal of Environmental Research and Public Health* 16, 1147. https://doi.org/10.3390/ijerph16071147
- Markkola J.A. and Karvonen R.T. (2020). Changing environmental conditions and structure of a breeding population of the threatened Lesser Whitefronted Goose (Anser erythropus L.). Ornis Fennica 97, 113–130. https:// doi.org/10.51812/of.133971
- Marolla F., Aarvak T., Hamel S., Ims R.A., Kéry M., Mellard J.P. et al. (2023). Life-cycle analysis of an endangered migratory goose to assess the impact of conservation actions on population recovery. *Biological Conservation* 281, 110028. https://doi.org/10.1016/j.biocon.2023.110028
- Molino del D., Seth V.J., Gyllenstrand N., Widemo F., Liljebäck N., Svensson M. et al. (2020). Population genomics reveals lack of greater white-fronted introgression into the Swedish lesser white-fronted goose. *Scientific Reports* 10, 18347. https://doi.org/10.1038/s41598-020-75315-y
- Morozov V.V. (1995). Status, distribution and trends of the Lesser Whitefronted Goose (*Anser erythropus*) population in Russia. *Casarca* 1, 131–144. (In Russian with English summary)
- Morozov V.V. (2021). Piskulka Anser erythropus (Linnaeus, 1758). In Amirkhanov A.M. et al. (eds), Red Data Book of Russian Federation. Moscow: FGBU VNII Ecologia Publishing House, pp. 561–563. (In Russian)
- Morozov V.V. and Kalyakin V.N. (1997). Lesser White-fronted Goose (Anser erythropus) in Southern Yamal: Retrospective analysis of population changes. Casarca 3, 175–191. (In Russian with English summary)
- Morozov V. and Syroechkovski Jr E. (2002). Lesser White-fronted Goose on the verge of the Millenium. Casarca 8, 233–276. (In Russian with English summary)
- Øien I. and Aarvak T. (2008). Lesser White-fronted Goose in Norway. Status of knowledge and proposal for National Action Plan. NOF/BirdLife Norway report 3-2008. Trondheim: Birdlife Norway.
- Øien I. and Aarvak T. (2009). The effect of Red Fox culling in the core breeding area for Fennoscandian Lesser White-fronted Geese in 2008. In Tolvanen P.,
 Øien I.J. and Ruokolainen K. (eds), *Final Report of the EU Life-Nature Project* 2005–2009. WWF Finland Report No27, NOF Rapportserie Report 1-2009. Trondheim: Birdlife Norway, pp. 81–82.
- Pokrovskaya O., Sokolova N., Erich D., Gilg O., Sokolov V. and Sokolov A. (2023). Globally threatened Lesser White-fronted Goose Anser erythropus nesting in association with Peregrine Falcons Falco peregrinus in southern Yamal, Russia. Wildfowl 73: 238–249.
- Reed E.T., Gauthier G. and Giroux J.F. (2004). Effects of spring conditions on breeding propensity of greater snow goose females. *Animal Biodiversity and Conservation* 27, doi:10.32800/abc.2004.27.0035
- Romanov A.A. (2003). New breeding areas of lesser white-fronted geese on the Putorana Plateau. *Casarca* **9**, 139–153 (In Russian with English summary)

- Romanov A.A. (2004). Vertebrate Fauna of the Putorana Plateau. Moscow: Nauka. (In Russian with English summary)
- Romanov A.A. and Pospelov I.N. (2010). Intracontinental spatial connections of Lesser White-Fronted Geese (*Anser erythropus*) from mountain subarctic regions of the Central Palearctic. *Russian Journal Ecology* **41**, 63–66. (in Russian with English summary).
- Ruokonen M., Kvist L., Tegelström H. and Lumme J. (2000). Goose hybrids, captive breeding and restocking of the Fennoscandian populations of the Lesser Whitefronted Goose (Anser erythropus). *Conservation Genetics* 1, 277–283.
- Rozenfeld S.B., Kirtaev G.V., Rogova N.V. and Soloviev M.Y. (2019). Results of an aerial survey of the western population of *Anser erythropus* (Anserini) in autumn migration in Russia 2017. *Nature Conservation Research* 4. https://doi.org/10.24189/ncr.2019.003
- Rozenfeld S.B. (2009). Feeding of brants and geese in the Russian Arctic. Moscow: KMK scientific publications association. 236 p. (in Russian).
- Romanov A.A. (2015). Avifauna of the Putorana Plateau. Moscow: Russian Society for the Study of Birds named after M.A. Menzbir. 196 p. (in Russian).
- Schreven K.H.T., Stolz C., Madsen J. and Nolet B.A. (2021). Nesting attempts and success of Arctic-breeding geese can be derived with high precision from accelerometry and GPS-tracking. *Animal Biotelemetry* 9, 25. https://doi. org/10.1186/s40317-021-00249-9
- Sedinger J.S., Chelgren N.D., Ward D.H. and Lindberg M.S. (2008). Fidelity and breeding probability related to population density and individual quality in black brent geese *Branta bernicla nigricans. Journal* of Animal Ecology 77, 702–712. https://doi.org/10.1111/j.1365-2656. 2008.01403.x
- Solovieva D. and Vartanyan S. (2011). Lesser White-Fronted Goose Anser erythropus: good news about the breeding population in west Chukotka, Russia. Wildfowl 61, 108–118.
- Tian H., Solovyeva D., Danilov G., Vartanyan S., Wen L., Lei J. et al. (2021). Combining modern tracking data and historical records improves understanding of the summer habitats of the Eastern Lesser White-fronted Goose Anser erythropus. Ecology and Evolution 11, 4126–4139. https://doi.org/10. 1002/ece3.7310
- Wang Y., Damba I., Zhao Q., Xie Y., Deng X., Ga R. et al. (2021) Organizing a juvenile ratio monitoring programme for 10 key waterbird species in the Yangtze River floodplain: analysis and proposals. Avian Research 12, 72. https://doi.org/10.1186/s40657-021-00309-x
- Wang X., Fox A.D., Cong P., Barter M. and Cao L. (2012). Changes in the distribution and abundance of wintering Lesser White-fronted Geese Anser erythropus in eastern China. Bird Conservation International 22, 128–134. https://doi.org/10.1017/S095927091100030X
- Xu F. and Si Y. (2019). The frost wave hypothesis: How the environment drives autumn departure of migratory waterfowl. *Ecological Indicators* 101, 1018–1025. https://doi.org/10.1016/j.ecolind.2019.02.024
- Yurkovskaya T.K. (2011). Vegetation map. In Shoba S.A. (ed.), Atlas pochv RF. Available at https://soilatlas.ru/karta-rastitelnosti. (In Russian).